

# Sequential bioethanol and methane production from municipal solid waste: An integrated biorefinery strategy towards cost-effectiveness

Antonio David Moreno<sup>a</sup>, José Antonio Magdalena<sup>b</sup>, José Miguel Oliva<sup>a</sup>, Silvia Greses<sup>b</sup>, Caterina Coll Lozano<sup>c</sup>, Marcos Latorre-Sánchez<sup>c</sup>, María José Negro<sup>a</sup>, Ana Susmozas<sup>a</sup>, Raquel Iglesias<sup>a</sup>, Mercedes Llamas<sup>b</sup>, Elia Tomás-Pejó<sup>b</sup>, Cristina González-Fernández<sup>b\*</sup>

<sup>a</sup>CIEMAT, Advanced Biofuels and Bioproducts Unit, Avda. Complutense 40, 28040 Madrid, Spain

<sup>b</sup>Biotechnological Processes Unit, IMDEA Energy, Avda. Ramón de la Sagra 3, 28935 Móstoles, Madrid (Spain)

<sup>c</sup>IMECAL S.L., Carretera de Carlet 74, 46250 L'Alcúdia, Spain

\*Corresponding author: [cristina.gonzalez@imdea.org](mailto:cristina.gonzalez@imdea.org); Telephone number: +34 917371127

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4

5 **Antonio David Moreno<sup>a</sup>, José Antonio Magdalena<sup>b</sup>, José Miguel Oliva<sup>a</sup>, Silvia**  
6 **Greses<sup>b</sup>, Caterina Coll Lozano<sup>c</sup>, Marcos Latorre-Sánchez<sup>c</sup>, María José Negro<sup>a</sup>, Ana**  
7 **Susmozas<sup>a</sup>, Raquel Iglesias<sup>a</sup>, Mercedes Llamas<sup>b</sup>, Elia Tomás-Pejó<sup>b</sup>, Cristina**  
8 **González-Fernández<sup>b\*</sup>**

9

10 *<sup>a</sup>CIEMAT, Advanced Biofuels and Bioproducts Unit, Avda. Complutense 40, 28040*  
11 *Madrid, Spain (email: david.moreno@ciemat.es, josemiguel.oliva@ciemat.es,*  
12 *mariajose.negro@ciemat.es, anaisabel.susmozas@ciemat.es,*  
13 *raquel.iglesias@ciemat.es)*

14 *<sup>b</sup>Biotechnological Processes Unit, IMDEA Energy, Avda. Ramón de la Sagra 3, 28935*  
15 *Móstoles, Madrid (Spain) (email: joseantonio.magdalena@imdea.org,*  
16 *silvia.greses@imdea.org, mercedes.llamas@imdea.org, elia.tomas@imdea.org,*  
17 *crisrina.gonzalez@imdea.org)*

18 *<sup>c</sup>IMECAL S.L., Carretera de Carlet 74, 46250 L'Alcúdia, Spain (email:*  
19 *caterina@imecal.com, marcos@imecal.com)*

20

21 *\*Corresponding author: cristina.gonzalez@imdea.org; Telephone number: +34*  
22 *917371127*

23

24

25 **Abstract**

26 The organic fraction of municipal waste (OFMW), source-sorted (SS-OFMW) and non-  
27 sorted (NS-OFMW), was used as raw material for the sequential production of  
28 bioethanol and biogas. Non-isothermal and simultaneous saccharification and  
29 fermentation (NSSF) resulted in maximum ethanol concentrations of 51 g/L and 26 g/L  
30 for SS-OFMW and NS-OFMW samples, showing overall process yields of up to 80%  
31 and 59%, respectively, even without subjecting substrate to hydrothermal pretreatment.  
32 Subsequently, the solid residues resulting from the fermentation were further subjected  
33 to anaerobic digestion (AD), showing a methanogenic potential of  $384 \pm 6$  mL CH<sub>4</sub>/g of  
34 volatile solids (VS<sub>in</sub>) and  $322 \pm 3$  mL CH<sub>4</sub>/g VS<sub>in</sub>, respectively. These methane yields  
35 were similar or even higher to those obtained when using non-fermented OFMW  
36 substrates (SS-OFMW:  $380 \pm 18$  mL CH<sub>4</sub>/g VS<sub>in</sub> and NS-OFMW:  $239 \pm 4$  mL CH<sub>4</sub>/g  
37 VS<sub>in</sub>), highlighting NSSF as a beneficial step to enhance methane yields during AD.  
38 Overall, bioconversion of OFMW would benefit from coupling bioethanol and biogas  
39 production since the biogas produced might be further employed as bioenergy source to  
40 compensate operational costs.

41

42 **Keywords:** anaerobic digestion; bioethanol; fermentation; methane; organic fraction  
43 municipal waste

44

45 **Abbreviation list**

46 AD            Anaerobic digestion  
47 BMP          Biochemical methane potential  
48 EU            European Union

49	GC	Gas chromatography
50	GHG	Greenhouse gases
51	MSW	Municipal solid waste
52	NSSF	Non-isothermal and simultaneous saccharification and fermentation
53	OFMW	Organic fraction of municipal waste
54	NS-OFMW	Non-sorted organic fraction of municipal waste
55	SS-OFMW	Source-sorted organic fraction of municipal waste
56	TS	Total solids
57	VS	Volatile solids

58

## 59 **1. Introduction**

60 Although many countries are trying to limit production of municipal solid waste (MSW)  
61 through implementation of policies for waste reduction, increasingly large amounts of  
62 these residues are produced. Today in Europe, each person generates an average of 475  
63 kg of MSW every year (Eurostat, 2018). According to the European Environment  
64 Agency (EEA), biowaste fraction accounts for more than 34% of the MSW generated in  
65 the European Union (EU), producing near 100 million tons (86 million tons in 2017) of  
66 biomass per year (European Environment Agency, 2020). This great amount of organic  
67 fraction in MSW (OFMW) can be regarded as useful feedstock for the generation of a  
68 wide range of biobased products instead of being discarded as waste (Kwan et al.,  
69 2019). This approach is in line with the nowadays mandate for promoting, developing  
70 and implementing a sustainable circular economy.

71 In order to achieve the objectives set by the EU related to greenhouse gases  
72 (GHGs) reduction and renewable energy share by 2030, at least 14% of the final  
73 consumption of energy in the transport sector of every EU country will need to come

74 from renewable sources, with a total share of at least 3.5% of advanced biofuels  
75 (European Union, 2018). In this context, the production of biofuels from waste  
76 feedstocks has become of great interest. Municipal biowaste includes biodegradable  
77 garden waste, food and kitchen waste from households, offices, restaurants, wholesale,  
78 canteens, caterers and retail premises and comparable waste from food processing plants  
79 (European Commission, 2018). This OFMW mainly contains carbohydrates (30-40%),  
80 proteins (5-15%) and lipids (10-15%) in terms of dry weight. Thus, bioethanol is an  
81 interesting biofuel that can be produced from this raw material due to its high  
82 carbohydrate content. This biofuel is the most globally used with a total production of  
83 85 billion liters in 2017 (World Bioenergy Association, 2019). Bioethanol is mainly  
84 obtained from sugar or starch containing products, such as sugar beet, corn and wheat  
85 crops. However, the use of lignocellulosic sources for bioethanol production, including  
86 the OFMW, would be preferred due to the competition with food and the concerns  
87 about ecological systems (Moreno et al., 2017).

88         After bioethanol production from OFMW, a significant amount of organic  
89 material from the feedstock still remains and can be transformed into biogas by  
90 anaerobic digestion (AD). Since only cellulose, starch and some dissolved  
91 carbohydrates are converted to bioethanol, a great fraction of the biowaste energy  
92 content is still present in the effluent of the fermentation process (mainly non-fermented  
93 carbohydrates, lipids and proteins). This makes the effluents from the fermentation  
94 process a suitable substrate for biogas production.

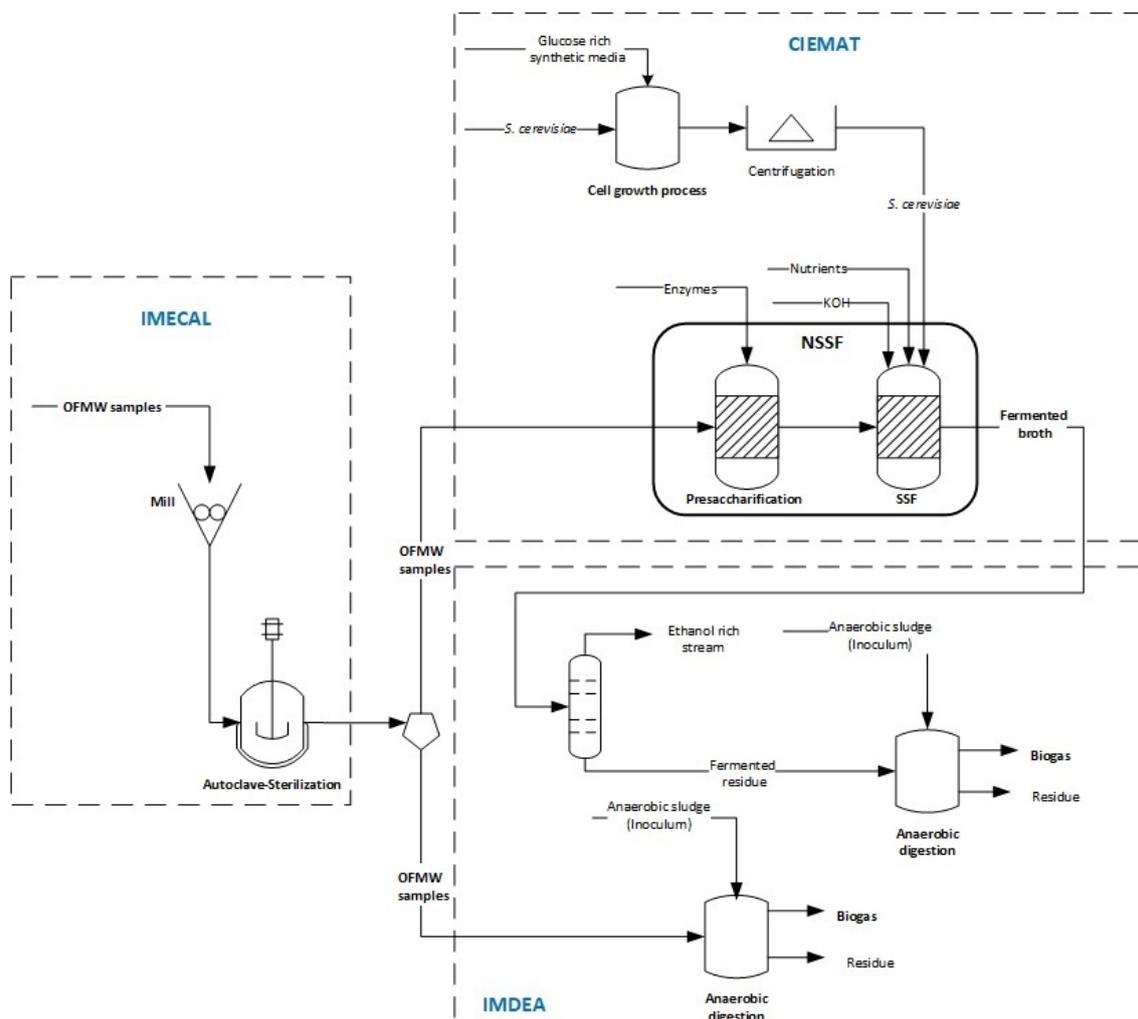
95         Co-production of bioethanol and biogas from different biomass has been studied  
96 previously (Karimi and Karimi, 2018; Park et al., 2012; Patel, 2017). These studies  
97 showed that biogas production in the corresponding bioethanol production facilities can  
98 contribute to the energy requirements of ethanol production as well as to increase the

99 energy yields from substrates, since biogas can be placed in local markets as biofuel or  
100 electricity.

101         Despite the benefits of sequential production of biofuels, improving the  
102 availability of the organic fraction in urban biowaste is one of the key elements for a  
103 profitable use of this substrate. The composition of OFMW that ends in a treatment  
104 plant is very diverse (Nielfa et al., 2015). The factors that influence this composition  
105 mainly include the waste collection system that has been used, the region where the  
106 MSW was generated and the climate variations during the year. The first factor  
107 influences the amount of inert materials (glass, plastics, textiles, sand, etc.) that will be  
108 contained in the biowaste fraction. This is a very important issue because the presence  
109 of a high content of inert materials causes many technological problems in plant  
110 facilities and reduces the efficiency of downstream valorization processes. In this  
111 manner, pretreatments such as mechanical sorting of the OFMW are essential to achieve  
112 the desired conversion efficiency of this complex feedstock into biofuels (bioethanol  
113 and biogas) by means of fermentative processes.

114         This investigation was designed to evaluate an integrated biorefinery strategy for  
115 the sequential production of bioethanol and biogas using OFMW as substrate (Figure 1).  
116 For such a purpose, two set of samples (source-sorted (SS) and non-sorted (NS)  
117 OFMW) collected at different seasons and locations were used as substrates with the  
118 aim of minimizing the effects related to compositional variations. With this purpose,  
119 bioethanol and biogas yields were determined for both types of samples. Special  
120 attention was given to the hydrolysis required for making the carbohydrate fraction  
121 available prior to the first fermentative process. At this point, it should be highlighted  
122 that while other lignocellulosic feedstocks have been extensively studied for the

123 production of advanced biofuels, the use of OFMW is envisaged as a challenging  
 124 substrate due to its complex and heterogeneous nature.



125  
 126 **Figure 1.** Experimental scheme followed in the present work for the sequential production of bioethanol  
 127 and biogas from source-sorted (SS-OFMW) and non-sorted (NS-OFMW) biowaste. NSSF, non-isothermal  
 128 and simultaneous saccharification and fermentation; SSF, simultaneous saccharification and fermentation.

129

## 130 2. Materials and Methods

### 131 2.1. Substrates

132 Taking into account the main factors that can affect OFMW composition, 16  
 133 representative OFMW samples were selected to be chemically characterized from South  
 134 (Spain, 8 samples) and North (United Kingdom, 8 samples) of Europe. Out of the eight  
 135 samples considered in each case, four came from separate collection systems

136 (considering both food waste and garden waste) and the other four from non-separate  
137 collection systems. Moreover, in order to address seasonal variability, samples were  
138 collected during spring, summer, autumn and winter. The collected samples (20 kg  
139 each) came from different industrial MSW treatment plants after mechanical  
140 pretreatments. Thereafter, the remaining inert materials content (glass, plastics, stones,  
141 textiles, etc.) in OFMW samples were removed manually at PERSEO Biorefinery Plant  
142 at IMECAL (L'Alcúdia, Spain) and then, samples were treated in a pilot plant mill in  
143 order to get a homogeneous feedstock. For this aim, an internal sampling protocol from  
144 IMECAL company was followed to avoid discrepancies in the OFMW samples  
145 collection. Samples were then sterilized at 121°C for 1h in autoclave to avoid sugar loss  
146 during transportation and storage (-20°C).

#### 147 *2.2. Fermentative microorganism*

148 *Saccharomyces cerevisiae* (Lesaffre, France) was used as fermentative microorganism.  
149 Active cultures of this yeast were obtained by growing cells on glucose synthetic  
150 medium (20 g/L glucose, 2 g/L yeast extract, 1 g/L NH<sub>4</sub>Cl, 1 g/L KH<sub>2</sub>PO<sub>4</sub>, and 0.3 g/L  
151 SO<sub>4</sub>Mg·7H<sub>2</sub>O) at 35°C, pH 5 and 150 rpm. After 16 h, cells were harvested by  
152 centrifugation at 5000 g for 5 min, washed once with 0.9% NaCl and diluted with  
153 distilled water to obtain the desired inoculum concentration.

#### 154 *2.3. Non-isothermal and simultaneous saccharification and fermentation of the OFMW*

155 Non-isothermal and simultaneous saccharification and fermentation (NSSF) processes  
156 were performed in triplicate in 250-mL shake flasks using 100 g of 20% (w/w) OFMW  
157 substrates. First, presaccharification process was performed at 50°C and pH 5 for 24 h  
158 (during hydrolysability tests, this saccharification step was extended up to 48 h). A  
159 tailor-made preparation containing both cellulases and amylases (kindly provided by

160 Novozymes, Denmark) was used for saccharification of OFMW samples (López-Gómez  
161 et al., 2019). The enzyme dose was chosen according to the supplier instructions, using  
162 the same mg of enzyme per g of glucan for comparison purposes. After  
163 presaccharification, temperature was reduced to 35°C, pH adjusted to 5.5 with 10 M  
164 KOH and the resulting hydrolysate was supplemented with the corresponding nutrients  
165 (the same components of the aforementioned rich medium without glucose; Section 2.2)  
166 and 1 g/L (dry weight) of *S. cerevisiae*. Flasks were then incubated in an orbital shaker  
167 at 150 rpm for 48-72 h.

#### 168 *2.4. Biochemical methane potential and sludge employed as seed inoculum*

169 To evaluate the methane potential of the residues (raw biowaste and fermented  
170 biowaste), anaerobic batch tests were conducted. Batch tests were run in triplicate at  
171 35°C and 150 rpm in serum glass bottles of total volume of 120 mL and working  
172 volume of 70 mL. In all biochemical methane potentials (BMPs), substrate to inoculum  
173 ratio was  $VS_{\text{substrate}}/VS_{\text{inoculum}} = 0.5$  (where VS refers to volatile solids). pH was adjusted  
174 to 7.5 at the beginning of the assay but not further controlled. 0.5 g CaCO<sub>3</sub>/L was  
175 supplied to each bottle to buffer the system and prevent pH changes. Bottles were  
176 flushed with helium to ensure anaerobic conditions. The biogas volume was calculated  
177 by measuring the pressure of the bottle's headspace. For the determination of  
178 endogenous methane production, blanks containing only anaerobic sludge were run.  
179 The overall methane production was calculated by subtracting the blank productions  
180 measured in each sample.

181 The anaerobic sludge used as inoculum was kindly provided by the wastewater  
182 treatment plant of Valladolid (Spain). Total solids (TS) and volatile solids (VS) of the  
183 inoculum were 23.7 g/L and 17.1 g/L, respectively.

184 *2.5. Analytical methods*

185 The chemical composition of OFMW samples after mechanical pretreatment and  
186 homogenization was analyzed in terms of carbohydrates (total glucans, starch, xylans,  
187 and other carbohydrates), protein content, ash content, and moisture using the  
188 Laboratory Analytical Procedures (LAP) for biomass analysis provided by the National  
189 Renewable Energies Laboratory (NREL, Golden, CO, USA) (Sluiter et al., 2008). In  
190 brief, samples were air-dried at 40°C until constant moisture (ca. 10%), and milled using  
191 a centrifugal mill (Retsch ZM200, Retsch, Ins., Haan, Germany) to 1 mm particle size.  
192 Total glucans, xylans and other carbohydrates were determined after a two-step acid  
193 hydrolysis using 1) 72% (w/w) sulfuric acid at 30°C for 60 minutes and 2) 4% (w/w)  
194 sulfuric acid at 121°C for 60 min. The resulting liquid fraction was analyzed by high  
195 performance liquid chromatography (HPLC) to determine sugar concentrations.  
196 Glucose, xylose, galactose, arabinose, and mannose were analyzed by a Waters HPLC  
197 system (Milford, MA, USA) equipped with a refractive index detector (model 2414)  
198 and a Transgenomic CARBOsep CHO-782 column (Omaha, NE, USA) working at  
199 70°C with ultrapure water (0.6 mL/min) as mobile phase.

200 To determine starch content, samples were milled to 0.5 mm in a Foss Cyclotec  
201 1093. This carbohydrate was measured using the Total Starch Assay Kit (Megazyme,  
202 Ireland) based on the use of thermostable  $\alpha$ -amylase and amyloglucosidase.

203 Total carbohydrate content was determined by the phenol sulphate method  
204 (Dubois et al., 1956). Nitrogen and protein content were estimated by the Kjeldahl  
205 method using a Tecator digester and Foss Tecator Kjeltex 8200 Auto Distillation Unit,  
206 considering a nitrogen-protein conversion factor of 6.25.

207 For TS measurements, the samples were placed on a crucible, weighted in a  
208 balance (Sartorius TE64, Germany) and subsequently introduced in an oven in  
209 accordance with APHA standard methods (Eaton et al., 2005). To determine the VS  
210 content, the sample resulting from the TS procedure was incinerated at 550°C for 5 h.  
211 The decrease in crucible weight represents the VS contained in the sample.

212 Ethanol was analyzed by gas chromatography (GC) using a HP 5890 Series II  
213 with an Agilent 6890 series injector. The system was equipped with a flame ionization  
214 detector and a Carbowax 20 M column, operating at 85°C. Injector and detector  
215 temperature were kept constant at 150°C.

216 Methane content in the biogas was determined by GC coupled with a thermal  
217 conductivity detector (Clarus 580 GC, PerkinElmer, USA) and equipped with an HSN6-  
218 60/80 Sulfinert P packed column (7' x 1/8" O.D.) and a MS13X4-09SF2 40/60 P packed  
219 column (9'x 1/8" O.D.) (PerkinElmer).

220

### 221 **3. Results and discussion**

#### 222 *3.1. Characterization of the OFMW*

223 To minimize the effects promoted by substrate variability, this investigation evaluated 8  
224 separately collected (SS-OFMW) and 8 non-separately collected (NS-OFMW) samples  
225 during different seasons from South (Spain) and North (United Kingdom) Europe.

226 Table 1 shows the chemical composition of OFMW substrates in terms of average and  
227 range values.

228 Regardless of the collection system, country or season, glucan was the main  
229 component of biowaste and makes this substrate an attractive alternative for  
230 fermentation-based processes. The main carbohydrates contained in this glucan fraction

231 were starch and other glucans (including cellulose). These components were shown to  
 232 be highly influenced by the collection system when compared to any other  
 233 macromolecule (Table 1). The average glucan content in SS-OFMW substrates was  
 234 50% higher than NS-OFMW, with maximum glucan content of about 60% and 35% of  
 235 the total dry matter for SS-OFMW and NS-OFMW, respectively. In addition to the  
 236 higher glucan content, SS-OFMW also exhibited a higher starch content when  
 237 compared to other glucans. Starch content in SS-OFMW ranged 23-43%, while NS-  
 238 OFMW substrates showed a starch content of 6-17%. This might be attributed to a  
 239 higher percentage of food-derived residues in SS-OFMW. Biowaste considers food  
 240 waste (starch-based biomass) and garden waste (cellulose-based biomass) for selective  
 241 sorting. Additionally, different collection systems can be used (i.e. garden waste and  
 242 food waste can be collected either separately or together) (Seyring et al., 2015), which  
 243 ultimately influences the proportions of these two wastes. In this work, systems  
 244 collecting both food waste and garden waste were used for the recovery of SS-OFMW  
 245 substrates. Nonetheless, the high ratio of starch vs. other glucans and the high starch  
 246 content in the studied SS-OFMW samples suggest a higher proportion of food-derived  
 247 waste in comparison to garden waste. In general, the amount of food waste discarded  
 248 yearly is higher than the garden waste. According to EEA, about 60% of the total  
 249 production of biowaste is food waste, while 30% is garden waste (European  
 250 Environment Agency, 2020).

251 **Table 1.** Minimum and maximum content for each component of the separately (SS-OFMW) and non-  
 252 separately (NS-OFMW) collected substrates.

Component	SS-OFMW		NS-OFMW	
	% (w/w)		% (w/w)	
<b>Total glucans</b>	33.9 - 61.3	(44.6 ± 9.7)	26.4 - 33.6	(29.8 ± 2.6)
<i>Starch</i>	23.5 - 43.1	(33.8 ± 7.5)	5.8 - 16.7	(10.3 ± 3.9)
<i>Other glucans</i>	4.8 - 18.2	(10.9 ± 5.2)	11.7 - 26.4	(19.5 ± 5.3)
<b>Xylans</b>	1.1 - 3.2	(1.7 ± 0.7)	2.6 - 4.4	(3.4 ± 0.7)
<b>Other carbohydrates</b>	2.4 - 5.3	(3.6 ± 1.2)	2.4 - 5.5	(3.3 ± 1.0)
<b>Protein</b>	12.4 - 17.3	(14.8 ± 1.7)	8.0 - 12.0	(9.7 ± 1.4)
<b>Ash</b>	6.2 - 9.7	(7.7 ± 1.3)	22.3 - 31.0	(27.1 ± 3.0)
<b>Moisture</b>	61.2 - 80.5	(69.7 ± 6.5)	49.3 - 59.7	(53.7 ± 3.8)

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Lipids and acid insoluble solids were not determined. Mean values  $\pm$  SD from n = 16 (8 substrates x 2 technical replicates for separately and 8 substrates x 2 technical replicates for non-separately collected) are listed in brackets.

253

254 Proteins and ash are also major components of OFMW substrates and were  
255 influenced by the waste sorting system as well (Table 1). In this context, SS-OFMW  
256 showed about 12-17% protein content and 6-10% ash content, while NS-OFMW  
257 exhibited 8-12% protein content and 22-31% ash content. It should be noted the high  
258 ash content in NS-OFMW ( $27.1 \pm 3.0$  g per 100 g of substrate). This result may be  
259 attributed to the presence of higher inert materials in NS-OFMW, even after the sorting  
260 processes (Alessi et al., 2020). The presence of a higher ash content can negatively  
261 influence the total glucan content in these samples, which might be the reason for the  
262 lower glucan concentrations in NS-OFMW.

263 Overall, the raw material was suitable to conduct the two-stage process. Given  
264 the glucan-rich nature of the biowaste, ethanol production seemed to be an excellent  
265 choice for valorizing this residual stream while the rest of the organic matter could be  
266 further used for biogas production purposes given the high organic matter content.

### 267 *3.2. Non-isothermal and Simultaneous Saccharification and Fermentation processes*

268 The high glucan content of OFMW (Table 1) makes this substrate attractive for several  
269 fermentation-based processes. Among them, the present work was focused on  
270 bioethanol production, since this chemical is leading current biofuel production  
271 worldwide (World Bioenergy Association, 2019) and has been recognized as an  
272 important building block to obtain industrially relevant alternatives (e.g. ethylene, ethyl  
273 acetate, acetaldehyde, etc.) (Posada et al., 2013). Bioethanol production from  
274 lignocellulosic substrates (including OFMW) requires ethanol titers of about 40 g/L to  
275 reduce the costs of the subsequent distillation step (Xiros et al., 2017). Higher ethanol

276 concentrations can be obtained by increasing substrate loadings. However, increased  
 277 substrate concentrations usually lead to lower process yields mainly due to a significant  
 278 reduction of the enzyme performance (Kristensen et al., 2009; Wang et al., 2011) and  
 279 the mixing conditions (Demichelis et al., 2017). In this sense, with the aim of evaluating  
 280 final saccharification yields, hydrolysability of OFMW samples was firstly assessed for  
 281 48 h at substrate concentrations as high as 20% (w/w). These tests showed average  
 282 glucose yields of about 60% and 50% for SS-OFMW and NS-OFMW, respectively  
 283 (Table 2).

284 **Table 2.** Glucose concentrations and saccharification yields obtained with separately (SS-OFMW)  
 285 collected and non-separately (NS-OFMW) collected samples

Substrate	Glucose (g/L)	Yield <sub>glucose</sub> (%) <sup>a</sup>
SS-OFMW	46 – 98	45 – 82 (63.7 ± 13.7)
NS-OFMW	30 – 47	43 – 56 (50.1 ± 4.6)

<sup>a</sup>Yields have been estimated as the percentage of the glucose released from the potential glucose contained in each substrate, subtracting the free glucose contained in the enzymatic preparation and substrates. Values representing average ± SD from n = 16 (8 substrates x 2 technical replicates for separately and 8 substrates x 2 technical replicates for non-separately collected) are indicated in bracket.

286

287 In general, SS-OFMW showed higher saccharification yields, reaching values as  
 288 high as 82%. This result is due to the fact that higher ratio of starch vs. other glucans  
 289 were observed for SS-OFMW samples, thus highlighting the good hydrolysability  
 290 potential of this substrate even when working at high substrate concentrations.  
 291 Compared to cellulose, enzymatic hydrolysis of starch polymer is relatively simple and  
 292 usually leads to higher hydrolysis yields (Salimi et al., 2019). In this context, samples  
 293 with higher ratios of starch vs. other glucans would exhibit higher saccharification  
 294 yields than other samples with similar total glucan content but lower ratios of these  
 295 carbohydrates. The higher saccharification yields, besides the higher total glucan  
 296 content in SS-OFMW, led to higher glucose concentrations when compared to NS-  
 297 OFMW (46-98 g/L vs. 30-47 g/L). It is remarkable that saccharification processes  
 298 resulted in a wide range of both glucose concentrations and yields. This is due to the

299 high variability found in OFMW substrates for the total glucan content and the ratio  
300 between starch and other glucans (Table 1). Glucose concentrations within these ranges  
301 have been reported previously at similar substrate loadings. López-Gómez et al. (2019)  
302 reported glucose concentrations of 37-55 g/L when subjecting OFMW to hydrolysis at  
303 20% (w/w) substrate loadings, while Demichelis et al. (2017) observed glucose  
304 concentrations of about 70 g/L when using food-derived biowaste at the same substrate  
305 loading.

306         Considering the results obtained during hydrolysability tests, bioethanol  
307 production from OFMW was evaluated at 20% (w/w) substrate loadings. In addition, a  
308 NSSF strategy with 24 h prehydrolysis was chosen for bioethanol production, since this  
309 configuration allows better integration of the different process steps (Moreno et al.,  
310 2017). NSSF of OFMW substrates led to ethanol concentrations of 24-51 g/L and 17-26  
311 g/L for SS-OFMW and NS-OFMW, respectively (Table 3). High overall process yields  
312 of up to 80% and 59% from the theoretical (and about 60% and 50% in average) were  
313 reached for these substrates. Similar to hydrolysability tests, NSSF processes exhibited  
314 high variation for ethanol concentrations and process yields within samples collected  
315 with the same sorting system. This variability correlates to the different nature and  
316 content of the glucan fraction as explained above (Table 1). Furthermore, it is important  
317 to highlight the high heterogeneity and the different origin of samples, which might  
318 have also affected both ethanol titers and yields.

319         In general, higher ethanol concentrations and yields were reached when using  
320 biowaste separated at source. This result is supported by the high saccharification yields  
321 observed for these samples and by the absence of microbial inhibition during NSSF  
322 processes. The higher starch content and the higher ratio between starch and other  
323 glucans of these substrates (Table 1) resulted in higher saccharification yields for SS-

324 OFMW (Table 2). On the other hand, no delays were found in ethanol production after  
 325 yeast inoculation during NSSF assays of SS-OFMW or NS-OFMW (Figure 2). Also,  
 326 final ethanol yields for both SS-OFMW and NS-OFMW were similar to the enzymatic  
 327 hydrolysis yields for these substrates (Table 2 and 3). These results can therefore  
 328 exclude any inhibitory effect of the microbial processes, evidencing enzymatic  
 329 hydrolysis as the major limitation during conversion of OFMW substrates and  
 330 highlighting the potential of OFMW as substrate for bioethanol production.

331 **Table 3.** Ethanol concentrations and saccharification yields obtained with separately (SS-OFMW) and non-  
 332 separately (NS-OFMW) collected samples

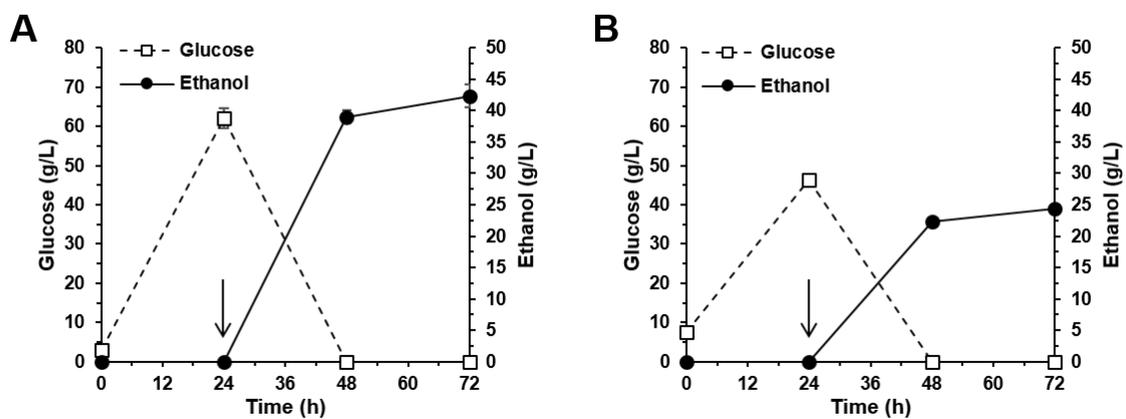
Substrate	Ethanol (g/L)	Yield <sub>etanol</sub> (%) <sup>a</sup>
SS-OFMW	24 – 52	46 – 80 (61.6 ± 12.7)
NS-OFMW	17 – 26	42 – 59 (52.2 ± 5.6)

<sup>a</sup>Ethanol yields have been estimated considering the maximum ethanol concentrations reached during NSSF processes and the potential glucose in samples (including glucose from enzyme preparation). Values representing average ± SD from n = 16 (8 substrates x 2 technical replicates for separately and 8 substrates x 2 technical replicates for non-separately collected) are indicated in bracket.

333

334 Final process yields as high as 60% were reached in average for SS-OFMW  
 335 substrates even though no thermochemical pretreatment process took place prior to  
 336 NSSF assays. These yields are similar to those obtained after subjecting NS-OFMW  
 337 substrates to thermal pretreatment. Ballesteros et al. (2010) reported final ethanol yields  
 338 of about 60% of the theoretical after subjecting OFMW to thermal pretreatment (active  
 339 hygienization). This pretreatment process requires to maintain 160°C for 30 min, which  
 340 influences the energy requirements and thus, the overall process costs. Also,  
 341 pretreatment processes generally release certain biomass degradations compounds that  
 342 are inhibitory for the hydrolytic enzymes and the fermentative microorganisms. For this  
 343 reason, pretreated materials usually require detoxification and/or highly tolerant  
 344 fermentative strains to trigger fermentation (Mahmoodi et al., 2018a, b). In this work,  
 345 the absence of a thermochemical pretreatment and the use of *S. cerevisiae* as  
 346 fermentative microorganism, which usually shows high tolerance to lignocellulose-

347 derived inhibitors (Moreno et al. 2019), resulted in no inhibition of the bioethanol  
 348 production process. Although further techno-economic assessments are needed,  
 349 avoiding the pretreatment and detoxification steps offers huge advantages for reducing  
 350 operational costs, since only pretreatment may represent up to 40% of the overall  
 351 process costs (Moreno et al., 2017). It is also interesting to note that ethanol  
 352 concentrations above 90% of the maximum observed were reached within the first 24 h  
 353 from inoculation (Figure 2). This also offers the possibility of shortening overall process  
 354 time and, consequently, reducing costs.



355  
 356 **Figure 2.** Time-course fermentations during non-isothermal and simultaneous saccharification and  
 357 fermentation (NSSF) of representative (A) separately (SS-OFMW) and (B) non-separately (NS-OFMW)  
 358 collected samples (Spain, spring). Arrows are indicative of the inoculation timing. Macromolecular  
 359 characterization of these particular samples can be found in Supplementary information.

360  
 361 *3.3. Methane yields attained via anaerobic digestion of the raw and fermented*  
 362 *substrates*

363 The remaining residue after ethanol distillation (stillage) is still a source of unconsumed  
 364 organic matter (mainly carbohydrates, proteins and lipids). Whereas glucose is the main  
 365 monomer employed for alcoholic fermentation, the rest of the carbohydrate fraction  
 366 (e.g. hemicelluloses), as well as proteins and lipids can be further digested for bioenergy  
 367 production purposes through AD. After ethanol production (fermented residue), the  
 368 characterization of samples revealed a slight increase of the rest of the macromolecules

369 with respect to carbohydrates. This was attributed to glucose consumption.

370 Macromolecular composition of the fermented residues is presented in Table 4.

371 The lower ash content in SS-OFMW and the percentages for the rest of the  
372 macromolecules were in accordance with values obtained in previous ethanol  
373 production assays as well as found in literature (Barampouti et al., 2019; Morales-Polo  
374 et al., 2018).

375 **Table 4.** Macromolecular composition of the fermented selective (SS-OFMW) and non-selective (NS-  
376 OFMW) collections employed in anaerobic digestion for methane production.

Component	SS-OFMW		NS-OFMW	
	Raw	Fermented	Raw	Fermented
Carbohydrates (%DW)	44 ± 4	24 ± 4	37 ± 3	24 ± 3
Proteins (%DW)	20 ± 3	25 ± 3	8 ± 2	16 ± 4
Lipids (%DW)	30 ± 9	43 ± 8	30 ± 8	33 ± 9
Ash (%DW)	5 ± 2	8 ± 1	25 ± 3	27 ± 2

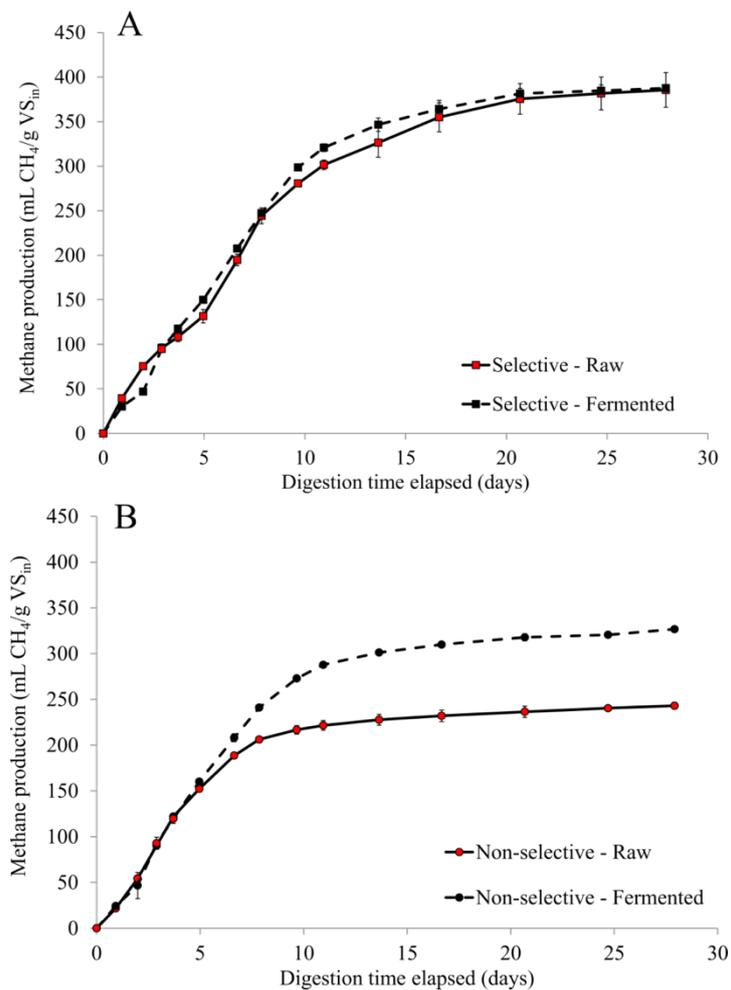
Values represent average ± SD.

377

378 In order to compare whether the AD process could contribute to further  
379 exploiting the unconsumed organic matter available in the fermented residue, non-  
380 fermented and fermented SS-OFMW and NS-OFMW collections were anaerobically  
381 digested for assessing their methanogenic potential.

382 The accumulated methane yields obtained can be observed in Figure 3. Methane  
383 potential attained by SS-OFMW were 380.9±18.4 mL CH<sub>4</sub>/g VS<sub>in</sub> and 384.6±6.5 mL  
384 CH<sub>4</sub>/g VS<sub>in</sub> for the raw and fermented residues, respectively, whereas NS-OFMW  
385 resulted in 239.9±4.2 mL CH<sub>4</sub>/g VS<sub>in</sub> and 321.7±2.9 mL CH<sub>4</sub>/g VS<sub>in</sub> (non-fermented  
386 and fermented residues, respectively). Residues coming from a selective collection (SS-  
387 OFMW) achieved higher methane yields than those attained in non-selective (NS-  
388 OFMW) batches (Figure 3). It seems that the SS-OFMW presented better properties  
389 than NS-OFMW when it comes to their anaerobic digestibility. Most probably, the high  
390 sorting efficiency decreased the presence of most recalcitrant fractions and improved

391 the organic fraction bioavailability, which was beneficial for methane yields. Likewise,  
 392 methane content in the raw NS-OFMW was 50% v/v CH<sub>4</sub> whereas higher values (60%  
 393 v/v CH<sub>4</sub>) were attained for the rest of the samples. With regard to the methane yield of  
 394 the non-fermented residues, the values obtained were in accordance to the values  
 395 available in literature (200-500 mL CH<sub>4</sub>/g VS<sub>in</sub> (Abudi et al., 2016; Allegue et al., 2020;  
 396 Bala et al., 2019; Nielfa et al., 2015b)). The wide range obtained in the different studies  
 397 in literature was attributed to the heterogeneity of the OFMW samples.



398

399 **Figure 3.** Methane yields of the fermented and raw fractions from (A) the selective collection (SS-OFMW)  
 400 and (B) the non-selective (NS-OFMW) collection.

401

402 In this work, biowaste was collected with two different sorting systems from

403 different locations, countries and seasons, leading to differences in the macromolecular

404 composition of OFMW. This wide variability of the chemical components of biowaste  
405 might results in variable methane yields. Alibardi and Cossu (2015) compared the  
406 methane production of three different fractions of OFMW with different  
407 macromolecular compositions (carbohydrate-rich, protein-rich and lipid-rich fractions).  
408 In that study, the methane production ranged from 400 mL to 600 mL CH<sub>4</sub>/g VS<sub>in</sub>  
409 depending on the assessed fraction. It should be highlighted that the methane yields  
410 obtained in the present study using the fermented residue as substrate were slightly  
411 higher when compared to other investigated substrates such as sugarcane bagasse,  
412 kitchen and garden wastes or algae (157-283 mL CH<sub>4</sub> / g VS<sub>in</sub>) (Karimi and Karimi,  
413 2018; Liang and McDonald, 2015; Park et al., 2012; Tian et al., 2013). This result  
414 showed the particular suitability of OFMW for the sequential production of bioethanol  
415 and biogas.

416           Compared to non-fermented residues, fermented residues resulted in similar or  
417 even higher methane yields (Figure 3). The increase in methane yields when using  
418 fermented residues has been also reported for other substrates. For instance, the  
419 fermented residue obtained from oat straw yielded 245 mL CH<sub>4</sub>/g VS<sub>in</sub>, while AD of the  
420 non-fermented raw material resulted in 201 mL CH<sub>4</sub>/g VS<sub>in</sub> (Dererie et al., 2011). This  
421 result might indicate that bioethanol production can influence substrate availability  
422 during the enzymatic saccharification and fermentation steps, which could make the  
423 organic matter more easily available for the anaerobic microbiome (i.e. the bioethanol  
424 production process may act as a pretreatment step in the digestion of the fermented  
425 residues). The high lipid content determined in the selective fermented residues might  
426 have prevented methane yield to increase more than the raw waste due to the toxicity  
427 exerted by these macromolecules. Lipids hydrolysis results in high amounts of long

428 chain fatty acids, which are detrimental for methanogens under concentrations higher  
429 than 40% in TS basis (Hu et al., 2018).”

430 Overall, the use of fermented OFMW and the raw residue gave as a result  
431 similar or even higher methane yields in batch mode. Therefore, coupling alcoholic  
432 fermentation and AD processes for bioethanol and energy generation might be regarded  
433 as a promising strategy to increase the cost-effectiveness of the process, even though  
434 these results still require confirmation in semi-continuous operation.

435

#### 436 **4. Conclusion**

437 SS-OFMW is a more preferable substrate for integrated biorefineries when compared to  
438 NS-OFMW due to i) its lower amount of inert materials, ii) the higher glucan content,  
439 and iii) the higher ratio between starch and other glucans. This raw material positively  
440 influences the bioethanol production by increasing the glucose available in this process.  
441 On the other hand, AD of fermentation residues results in similar and even higher  
442 methane yields than their raw counter partners regardless of biowaste collection type.  
443 This sequential strategy offers a more complete use of OFMW increased the carbon  
444 conversion yield of this substrate into energy.

445

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456

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