WATER LEACHING OF HERBACEOUS BIOMASS BALES TO REDUCE
SINTERING AND CORROSION
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ABSTRACT:
Two different leaching treatments (immersion and irrigation) were applied to the biomass bales
of two different herbaceous solid biofuels (barley straw and tall wheatgrass) in order to improve
their sintering and corrosion behaviour during combustion. Six immersion treatments (1 h with
and without stirring, and 5 h, 1 day and 2 days without stirring) and two irrigation treatments
with tap water (900 L and 1800 L) were tested. The bales made of barley straw were of high-
density 140 kg/m <sup>3</sup> ) whereas the density of tall wheatgrass bales was lower (80 kg/m <sup>3</sup> ). After
the leaching treatment, bales were dried naturally indoors. Sintering and corrosion tests along
with chemical elemental analyses, X-ray diffraction and SEM analyses were performed on the
lixiviated biomass.
An immersion treatment of 5 h reduced the sintering and corrosion potential of both types of

bales to an acceptable level. In high-density bales, the irrigation treatment (1800 L) provided

1	similar sintering and corrosion reductions than those obtained when the 5 h immersion
2	treatment was applied. However, in low-density bales, the immersion treatments provided much
3	better results than irrigation leaching systems, probably in connection with their different
4	degrees of compaction. The naturally drying process of the lixiviated bales can be considerably
5	long, particularly after immersing high-density bales. The drying period to reach a moisture
6	level of 20% w.b, was longer than 100 days for high-density bales and shorter than 20 days for
7	low-density bales.
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9	Keywords: biomass, leaching, corrosion, sintering, drying, chemical composition.
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14	1 INTRODUCTION
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16	Biomass is postulated as the main source of renewable energy to produce heat and electricity
17	where a deficit of wind or sun energies exists. However, good quality biomass, mainly of woody
18	origin, tends to increase the energy price [1] and thus other sources of biomass should also be
19	considered. According to their origin and physical and chemical properties, biomass derived
20	sources can be divided into: woody, herbaceous, agro-industrial, animal, human, and aquatic
21	biomass, as well as the organic fraction of urban or industrial wastes. Among all these origins
22	and sources, the herbaceous biomass might be the best solution due to its abundance and
23	relatively good properties such as their moisture content and calorific value.

25 Nevertheless, the herbaceous biomass might generate environmental problems derived from the

emission of particles, sulphur oxides and nitrogen oxides, as a consequence of its significant levels of ash, S and N, respectively. In general, herbaceous biomass is rich in alkaline compounds and is prone to cause technical problems in power plants such as fouling, slagging and corrosion when compared with the biomass of woody origin.

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6 Several solutions have been proposed to reduce corrosion and sinter formation in the thermo-7 chemical processes of herbaceous biomass, including biomass leaching [2-5], the use of mineral 8 additives, co-combustion with other fuels [5-9], decreasing the combustion temperature [5], 9 or even a combination of processes such as pyrolysis followed by combustion [10]. The 10 selection among all the available options will depend not only on the cost of the process but 11 also on the efficiency of the occurring chemical reactions.

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Laboratory tests on biomass leaching are very promising since they are highly effective in 13 14 modifying the composition of the fuel by removing water soluble compounds. It is due to the intimate contact that is reached between water and biomass, which can even be enhanced by 15 prolonging extraction times, applying agitation or reducing the particle size of the fuel. 16 However, with the exception of field leaching [2-3,9,11], most of these methods have been 17 barely studied, particularly when it comes to real field applications in pilot plants or on an 18 industrial scale. To the best of the authors' knowledge, industrial plants based on biomass water 19 leaching do not exist. 20

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One of the main problems of straw field leaching is the lack of rainwater needed to lixiviate alkaline elements, which might be a particularly severe issue in certain environments such as marginal lands suffering from water scarcity. Additional problems have been reported in the literature when using this field leaching treatment, such as sand contamination from soil, straw

losses derived from strong winds, or the loss of organic material as a result from biomass decomposition [11]. Moreover, biomass field drying could take long time and can be further extended when it rains. On the other hand, field leaching avoids the equipment costs needed when subjecting the fuel to a wetting treatment at a facility site [11] and might reduce the fertilisation inputs, as nutrients from leachates are able to return to soil.

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In this work, different straw leaching treatments are studied in a pilot plant facility in order to 7 reduce sintering and corrosion while avoiding the problems which may arise during field 8 leaching, such as rainwater deficits, material losses or biomass contamination with soil 9 particles. The procedures proposed could be performed in industrial facilities prepared for this 10 11 purpose, which could be located either in the field or at a power plant. Two different leaching treatments (immersion and irrigation) were studied under different conditions using low and 12 high density straw bales. In addition, the drying process of bales after being leached was 13 monitored indoors under natural ventilation. 14

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# 2 MATERIALS AND METHODS

3	2.1 Biomass and solid biofuels
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5	Two herbaceous biomass feedstocks were used: barley (Hordeum vulgare) straw from a
6	traditional crop and tall wheatgrass (Elytrigia elongata) from an energy crop. The whole of the
7	plant above ground (without roots,) of tall wheatgrass was used for baling.
8	
9	Biomass was harvested, pressed and tied into rectangular bales and air dried naturally. Physical
10	characteristics of the bales are shown in Table 1. High-density bales were reached in case of
11	barley straw (137 kg/m <sup>3</sup> ) and low-density bales (80 kg/m <sup>3</sup> ) were produced in the case of tall
12	wheatgrass. The bales were stored indoors under natural ventilation until treatment.
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14	Table 1: Bale characteristics
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18	2.2 Leaching treatments
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20	Two different leaching treatments (immersion and irrigation) were tested using tap water at
21	room temperature (Figure 1).
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23	During the intervencion treatment heles were submerced in a glastic container, which was
	During the immersion treatment, bales were submerged in a plastic container, which was

Four different immersion conditions were tested: 1 hour, 1 hour with stirring, 5 hours, 1 day

and 2 days. In case of the experiment that involved stirring, water was agitated with a shovel
for 1 minute and every 5 minutes.
Figure 1: Scheme of the two different bale leaching treatments
During the irrigation treatment, bales were rinsed using 900 or 1800 litres of tap water (Figure
1 and Table 2). Bales were placed on a pallet, which was located on the upper part of the
container used for water collection. Water was evenly sprayed on the upper surface of the bale
with a garden sprayer (head diameter of 5 cm), starting from one edge of the bale and ending at
the opposite edge until the surface of the bale was entirely showered. In the case of 900-litre
irrigation, every bale was irrigated twice with a total volume of 450 litres per irrigation. This
process lasted 90 min per each 450 litres. Consequently, in the case of 1800-litre irrigation,
each bale was irrigated four times with a total duration of 6 hours.
Table 2: Leaching treatments
Bales were weighed before and after being lixiviated. Water conductivity in the container was
monitored in order to control the leaching process. Biomass samples were extracted from bales
by using a hook and following the international standard for sampling solid biofuels (ISO
18135:2018). Three subsamples were taken from different positions of each bale before being

1	treated and three more subsamples were extracted after lixiviation. Subsamples were mixed to
2	form a combined sample for testing.
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6	2.3 Physical and chemical tests
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8	Biomass and ash analyses were carried out following current ISO standards for solid biofuels.
9	The analytical techniques and standards used are listed in Table 3.
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11	Biomass ashes, which were obtained at 550 °C, were acid digested in a microwave furnace
12	using HNO <sub>3</sub> , H <sub>2</sub> O <sub>2</sub> and HF as a first step and H <sub>3</sub> BO <sub>3</sub> as a second step before being further
13	analysed by ICP-OES (Thermo Jarrell Ash) in accordance with ISO 16967:2015.
14	
15	X-ray diffraction analyses were performed using a PANalytical X´pert-PRO diffractometer
16	operating at 45 KV and 45 mA. The mineral phases of the biomass ashes (obtained at 550 °C
17	following ISO 18122:2015) were identified by means of the Inorganic Cristal Structure
18	Database (ICSD, Fiz Karlsruhe, Germany). Semi-quantification was performed using Rietveld
19	analysis and results were normalised.
20	
21	Ash melting behaviour was determined according to ISO 21404:2020 [12] and using an optical
22	heating microscope (Hesse Instruments). The following three characteristic temperatures were
23	reported: deformation temperature (DT), hemisphere temperature (HT) and fluid temperature
24	(FT).
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1	Table 3: Standards and analytical methods
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5	In addition, the earth-alkaline to alkaline oxides ratio (I) was also used to predict ash melting
6	behaviour. This predictive index was previously developed by CIEMAT and tested successfully
7	in different solid biofuels [13]. Index I is calculated as follows:
8	$I = (CaO + MgO) / (K_2O + Na_2O)$
9	The ash melting behaviour improves as I increases. In general, ashes with I values higher than
10	2 should not present sintering risks at the normal operating temperatures of a biomass
11	combustor.
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15	2.4 Corrosion tests
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17	To predict how the lixiviation process could have affected the corrosion behaviour of the
18	biomass, a corrosion test at laboratory scale was performed on the ashes of the leached biomass.
19	Ashes were obtained from the biomass of the lixiviated bales at 550 °C following ISO
20	18122:2015 and placed on rectangular steel bases (20 mm x 10 mm, 2 mm thickness) as it is
21	shown in Figure 2. The selected steel was Fe-2.25Cr-1Mo. The steel coupons were previously
22	polished with sandpaper numbers 220, 500, and 1200, and cleaned with acetone. Ashes (50 mg
23	per sample) and steel coupons were placed on a sample holder which was able to accommodate
24	18 samples, and then introduced in a tubular furnace at 700 °C for 200 h with a continuous air
25	flow at 150 mL/min (Figure 2).

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4	Figure 2: Schematics of the corrosion test using a tubular furnace.
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6	Before SEM-EDX (Scanning Electron Microscopy - Energy Dispersive X-ray spectroscopy)
7	analysis, the steel coupons were embedded in resin, cut with a precision cut-off machine
8	(Struers Secotom-10) using an Al <sub>2</sub> O <sub>3</sub> disc and a mineral oil like refrigerant, polished with
9	sandpaper number 1200 and washed with hexane in an ultrasound bath.
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11	The steel thickness and the elemental composition of the metal/corrosion interface were
12	measured by SEM-EDX using a Zeiss instrument (Evo LS-10).
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16	2.5 Bale drying and meteorological conditions
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18	Wet bales were stored indoors under natural ventilation throughout most of the summer (from
19	May to September). Each bale was placed on a pallet for easy handling and transportation and
20	to speed up the drying process. The moisture content was determined as a difference between
21	final and initial mass using a digital hanging crane scale.
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23	Tests were performed in Soria (central-northern Spain). According to Köppen-Geiger
24	classification, the climate at the experimental site is considered to be temperate with a dry
25	season and mild summer (Cfb). During the bale drying study period (from May to September),

1	a mean temperature of 19 °C and maximum temperatures considerably above 30 °C were
2	registered, with a maximum absolute temperature of 35.3 °C. Mean ambient relative humidity
3	was 63%.
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## 1 3 RESULTS AND DISCUSSION

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#### **3.1** Physical and chemical composition of the baled biomass

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5 Table 4 shows the calorific values and the proximate and ultimate analysis of the baled biomass,

6 together with the ash fusibility temperatures and their ash chemical composition.

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Tall wheatgrass calorific value (mean net calorific value of 17.44 MJ/kg d.b.) is higher than for
barley (16.25 MJ/kg, d.b.) due to its higher C and H contents and its lower ash content (4.3%
vs. 7.1%).

11

In spite of the similarity of the chemical composition of both types of ashes, with comparable 12 SiO<sub>2</sub>, K<sub>2</sub>O and CaO levels, their deformation and fluid characteristic temperatures differed by 13 100 °C, probably in connection with their different Cl content, which is higher in barley straw 14 (0.62 wt.%) than in tall wheatgrass (0.18 wt.%). Consequently, the proportion of KCl is higher 15 in barley than in tall wheatgrass, as it was demonstrated by XRD analysis (see Section 3.2). In 16 contrast, barley exhibits lower S contents (0.06 %) than tall wheatgrass (0.10 wt.%). It is well 17 known that the melting temperature of KCl (775 °C) is lower than K<sub>2</sub>SO<sub>4</sub> (1045 °C). This could 18 be the reason why barley straw sinters and melts at lower temperatures than tall wheatgrass. 19 20 
**Table 4**: Characterisation of the barley and tall wheatgrass baled biomass.
 21 22 d.b.: dry basis; GCV: gross calorific value at constant volume; NCV: net calorific value at constant pressure; 23 ND: not determined 24

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2	<b>3.2</b> Effect of leaching on the chemical composition
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4	The concentrations of K, Cl and S of the baled biomass were analysed before and after applying
5 bale 6	the leaching treatments considered, and the results are shown in Figure 3 for tall wheatgrass as and Figure 4 for barley straw bales.
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10	Figure 3: S, Cl, and K concentrations on the baled tall wheatgrass biomass according to the
11	leaching treatment.
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15	Figure 4: S, Cl, and K concentrations on the baled barley straw biomass according to the
16	leaching treatment.
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18	K and Cl decreased notably with immersion time. The highest K, Cl, and S leaching efficiencies
19	were noticed when bales were submerged for 1 day. Longer immersion times (2 days) were not
20	able to leach higher amounts of K, Cl, and S from the baled biomass. In addition, the leaching
21	of these elements was also not affected positively by water stirring.
22	
23	After being immersed for 1 day, contents of 0.21% of K (74% of K reduction), 0.03% of Cl
24	(83% of Cl reduction), and 0.06% of S (39% of S reduction) were measured for the biomass of
25	tall wheatgrass, whereas the biomass of barley straw showed levels of 0.18% of K (86% of K

reduction), 0.06% of Cl (90% of Cl reduction), and 0.03% of S (58% of S reduction).

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Even though the content of S present in the baled biomass was clearly reduced after leaching,
these levels decreased to a lower extent than K and Cl, probably in connection with the lower
solubility of arcanite (K<sub>2</sub>SO<sub>4</sub>) when compared to sylvine (KCl). Specifically, the solubility of
KCl in cold water at 20 °C is 340 g/L, much higher than that of K<sub>2</sub>SO<sub>4</sub> (120 g/L).

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The efficiency of K leaching when tall wheatgrass bales were irrigated was lower than when bales were submerged, showing reductions of 21% vs. 74% for 1-day immersion. Additionally, when applying the irrigation treatment to this type of bale, the conductivity increase of the leached water, when compared to the tap water used for irrigation, was negligible. The lower K reductions observed when this type of bale were irrigated, when compared to being submerged, are probably connected to its low density, which does not foster the contact between the biomass fibres and the irrigation water.

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In contrast, irrigation seems to have affected the K content of barley straw bales to a higher extent than tall wheatgrass bales. K reductions of 49% and 67% were noticed when irrigating barley bales with 900 and 1800 L, respectively. The higher grade of compaction of barley straw bales seemed to favour this interaction by trapping water inside the bale, which could have caused a more intimate contact of the biomass with the water embedded in the bale. In any case, 5-hour to 1-day immersions seem to be the most efficient treatment to leach the soluble elements present in this type of biomass and bale.

23

Figures 5 and 6 show the mineral compounds of tall wheatgrass and barley straw ashes, respectively, according to the leaching treatment. As can be seen in these figures, the

1	concentration of KCl present in barley ashes was higher than that contained in tall wheatgrass,
2	whereas K <sub>2</sub> SO <sub>4</sub> levels were found to be higher in tall wheatgrass ashes, which is also related to
3	the different fusibility behaviour of the two biomasses studied (section 3.1).
4	
5	On the other hand, it should be mentioned that calcinite (KHCO <sub>3</sub> ) was only detected in the ashes
6	of tall wheatgrass bales before leaching. This compound was not detected once these bales were
7	leached, probably as a consequence of the low concentration in the original bale together with
8	its relatively high water solubility (220 g/L).
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12	Figure 5: Ash inorganic compounds of tall wheatgrass bales according to the leaching
13	treatment (XRD analysis)
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17	Figure 6: Ash inorganic compounds of barley straw bales according to the leaching treatment
18	(XRD analysis)
19	
20	In summary, the extent to which the water soluble compounds are leached will depend not only
21	on the solubility of the inorganic compounds contained in the biomass ash and the type and
22	conditions of the applied leaching treatment, but also on bale density.
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The predictive empirical index I (see section 2.3) is shown in Figure 7 as a function of the leaching treatment for the two types of bales studied. Higher indices were obtained for those samples characterised by lower K contents.

3.3 Effect of leaching on ash melting behaviour

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7 As mentioned in section 2.3, it was previously demonstrated that, in general, biomasses with earth-alkaline to alkaline oxides ratios higher than 2 should not present risk of sintering at 8 9 normal operating combustion temperatures (850 °C) [13]. Indices close to 0.5 were obtained when bales are not subjected to any type of leaching treatment (Figure 7). To reach the 10 recommended index levels to avoid sintering, a minimum immersion time of 5 hours was 11 necessary. As commented in section 3.2, immersion times longer than 1 day did not increase 12 13 K, Cl, or S leaching and, consequently, index I did not predict any improvement in the ash melting behaviour of the leached biomass either (Figure 7). Therefore, immersion times from 14 5 hours to 1 day would be the optimal leaching treatment to provide the lowest risk of ash 15 sintering for both types of bales. In addition and, also in accordance with the previous section, 16 index I predicts that water agitation does not favour ash melting behaviour. 17

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It is also worth mentioning that, an index I higher than 2 was obtained for barley straw bales irrigated with 1800 L, in connection with the noticeable K decrease observed when this type of bale was subjected to this leaching treatment.

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**Figure 7:** Earth-alkaline to alkaline oxides ratio (empirical index I) according to the leaching

1	treatment.
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3	The ash fusibility characteristic temperatures, analysed according to the ISO standard
4	21404:2020, are shown in Table 5 as a function of the leaching treatment. Since hemisphere
5	temperatures (HT) were not always detected, only deformation temperatures (DT) and fluid
6	temperatures (FT) were included in the table.
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8	<b>Table 5:</b> Ash melting characteristic temperatures (according to ISO 21404:2020) according to
9	the leaching treatment.
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11	DT: deformation temperature; FT: fluid temperature; ND: not determined
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15	The DT and FT of immersed samples tended to increase considerably as the immersion time
16	increased, rising from mean DTs of 810 °C (barley) and 900 °C (tall wheatgrass) up to values
17	well above 1150 °C after 1-immersion day. FT also varied extensively with immersion time for
18	the two biofuels tested; from initial mean values around 1100 °C to FTs above 1350 °C after 1
19	immersion day. It is important to note (Table 5) that, once again, water stirring does not seem
20	to affect positively the melting behaviour of tall wheatgrass or barley bales.
21	
22	Similarly to the results obtained when applying the predictive index I, water irrigation with 900
23	L did not improve the melting behaviour of tall wheatgrass ashes, and increased very slightly
24	the DT and FT of barley straw ashes (Table 5). In contrast, the irrigation of barley straw bales
25	with 1800 litres of water increased their ash melting temperatures remarkably, up to DT of 1160

<sup>1</sup> °C and FT of 1250 °C.

Therefore, the changes noticed in the ISO 21404:2020 characteristic temperatures with regard to the ash melting behaviour are in close agreement with the predictions obtained by the CIEMAT index I, and with the chemical composition of the biomass before and after bales were leached (section 3.2). 3.4 Effect of leaching on ash corrosion behaviour The reduction of metal thickness under the conditions detailed in section 2.4 (SEM analysis), was used as a measurement of the ash corrosive behaviour of the biomass of tall wheatgrass and barley straw. In this sense, Figure 8 shows the metal corrosion caused by the ashes obtained from the baled biomass before and after the application of the different leaching treatments studied. Figure 8: Metal corrosion (SEM analysis) according to the leaching treatment. All leaching systems led to a significant improvement on ash corrosion behaviour. Corrosion decreased remarkably with immersion time, from reductions in the metal thickness between 6-7% when bales were not treated until values 3% after bale immersions for 5 h or longer. 

Irrigating tall wheatgrass bales was not as effective as bale immersion in terms of improving ash corrosion behaviour, probably in connection with the poorest K leaching efficiencies obtained when irrigating this type of bale, as explained in section 3.2. In turn, the irrigation of barley straw bales with 900 and 1800 L proved to be more effective.

5

In general, there is a good correlation between the improvements in the ash corrosion (Figure 6 8) and sintering (Figure 7 and Table 4) behaviours with the K content of the baled biomass 7 (Figures 3 and 4). However, K biomass levels (Figures 3 and 4) suggested that the corrosion 8 behaviour of barley straw immersed bales should have been as good as that showed by tall 9 wheatgrass, i.e. the loss of metal thickness in leached bales (Figure 8) of barley straw should 10 11 have been as low as for tall wheatgrass if we take the similar potassium content into account (Figures 3 and 4). An explanation is that the effect of the biomass ash on corrosion is not directly 12 proportional to contents of inorganic elements such as potassium and many mechanisms, 13 chemical reactions and degree of interactions among the types of the inorganic compounds and 14 15 the metal interface exist.

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Figure 9 corroborates the above commented difference between barley straw and tall 17 wheatgrass, and shows how the ash/steel interface was reduced when bales were immersed for 18 1 day, for example. When bales were not lixiviated, the resulting corrosion products after 19 heating the ash/steel coupons at 700 °C for 200 h (section 2.4) formed a layer of about 400 µm. 20 The major corrosion product observed in this layer was iron oxide. The thickness of this 21 corrosive layer was reduced from about 400 µm to about 100 µm and 200 µm when tall 22 23 wheatgrass and barley straw bales were submerged for 1 day, respectively, showing the poorest corrosion reduction of barley. 24

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3	Figure 9: Ash/steel interfaces before and after 1-day bale immersions (SEM images).
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7	3.5 Effect of leaching on bale drying
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9	Bale drying is probably the most significant bottleneck to develop leaching systems on a
10	commercial scale, since it can be a considerably long process, which can greatly affect the
11	economic profitability of the entire approach.
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13	In this work, bales were dried indoors under natural ventilation and their moisture content was
14	monitored until reaching values below 20 %. Tall wheatgrass bales dried much faster than
15	barley straw bales, very probably in connection with the lower density of the former. In this
16	sense, the moisture content of tall wheatgrass bales was reduced below 20% (w.b.) after less
17	than 20 days (data not shown). In contrast, more than 100 days were needed for barley straw
18	bales (high-density bales) to reach those moisture levels (Figure 10). Therefore, bale
19	compaction seems to play a very important role when it comes to bale drying.
20	
21	As can be seen in Figure 10, the drying times for irrigated high-density bales were shorter
22	(between 100 and 120 days) than those needed when bales were submerged during the
23	immersion treatment. This is a consequence of the lower moistures reached by the bales after
24	irrigation (65%) in comparison with those attained after bale immersion (between 70% and

1	changes the access of water into the pores or spaces of the straw. The hydrostatic pressure in
2	the immersion process forces water into pores, which cause longer drying times. In contrast,
3	the negligible hydrostatic pressure in the irrigation process reduces the amount of water into
4	the pores decreasing drying times.
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8	Figure 10: Natural drying of high-density bales after being subjected to different leaching
9	treatments.
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11	Drying times as long as 140 days were needed to decrease the moisture content of bales, which
12	were previously submerged for 1 hour, below 20%. At the same time, it was clear that the longer
13	the immersion time the longer the drying time. Thus, drying times will be considerably longer
14	than 140 days to reach moistures below 20% in the bales submerged over five hours or more.
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### 1 4 CONCLUSIONS

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3 This work leads to the following conclusions:

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A complete leaching of all the water soluble elements in both high and low density herbaceous bales was reached after water immersions for 24 hours. Nevertheless, bale immersions for 5 h would be enough to reduce the sintering and corrosion of the baled biomass to acceptable levels.

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In the case of high-density bales, the irrigation treatment (1800 L) provided comparable sintering and corrosion reductions to those obtained with the 5 h immersion treatment. However, irrigation was not an effective leaching treatment for low-density bales, probably as a consequence of poorer contact between water and biomass fibers, which resulted in a less efficient lixiviation of alkaline compounds. Therefore, the degree of bale compaction appears to influence the lixiviation process and could be key in selecting the appropriate leaching treatment.

16

Furthermore, the leaching studies performed on the two types of herbaceous biomass led to the conclusion that the alkaline compounds present in barley straw, based on potassium chloride, are more readily leached than in tall wheatgrass, which are mainly composed of potassium sulphates. Therefore and, as expected, the water solubility of the different ash forming alkaline compounds affects leaching dynamics.

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The drying time of the lixiviated baled biomass depends on several factors such as bale density or the leaching system applied (immersion or irrigation). The drying time of high-density bales can be considerably long and therefore this aspect should be taken into account when designing

1	industrial processes. Immersion treatments in particular result in elevated biomass moisture
2	contents and therefore longer drying times. Under the conditions tested in this work, more than
3	100 days were necessary to naturally reduce the moisture of this type of bale below 20 wt.%
4	(wet basis). In turn, shorter periods (less than 20 days) were needed to reach the same moisture
5	levels when drying low-density bales.
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<u>\*</u>

	Tall wheatgrass	Barley straw
Weight (kg)	14	30
Size (cm)	100×38×46	125×50×35
Surface (m <sup>2</sup> )	0.46	0.75
Volume (m <sup>3</sup> )	0.17	0.22
Density (kg/m <sup>3</sup> )	80	137

Table 1: Bale characteristics

Т	able	e 2:	Leaching	treatments
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Tuble 2. Leading treatments					
1 hour					
	1 hour + stirring				
Immersion (into 900 litres)	5 hours				
	24 hours				
	48 hours				
Irrigation	900 litres (3 hours)				
inigation	1800 litres (6 hours)				

# Table 3: Standards and analytical methods

Property	Analytical technique	Standard
Sample preparation	Homogenisation, subsampling, milling	ISO 14780:2017
Moisture	Drying at 105 °C	ISO 18134-2:2017
Ash	Calcination at 550 °C	ISO 18122:2015
С, Н, N	Elemental analysis: IR and TCD (Leco, TruSpec)	ISO 16948:2015
S and Cl	Combustion bomb (Ika) + IC (Methrom)	ISO 16994:2016
Calorific value	Calorimetry (Ika, C5000)	ISO 18125:2017
Major-minor elements	MW digestion + ICP-OES (Thermo Fisher, Iris AP)	ISO 16967:2015
Inorganic compounds	X-ray diffraction (PANalytical, X´pert-PRO)	ISO 16968:2015
Ash melting behaviour	Optical heating microscopy (Hesse)	ISO 21404:2020

		Moan	ad	0/mad	Moan	ed	0/mad
Calorific valu	ues	mean	su	701 SU	mean	sa	70rsa
GCV	MJ/kg d.b.	18.78	0.08	0.45	17.53	0.29	1.7
NCV	MJ/kg d.b.	17.44	0.08	0.44	16.25	0.3	1.8
Proximate an	alysis						
Ash	wt.% d.b.	4.3	0.4	8.6	7.1	0.4	5.8
Ultimate ana	lysis						
Carbon	wt.% d.b.	44.6	0.4	8.6	44.3	0.6	1.4
Hydrogen	wt.% d.b.	6.1	0.05	0.84	5.9	0.1	1.3
Nitrogen	wt.% d.b.	1.0	0.2	19	0.42	0.09	21
Sulphur	wt.% d.b.	0.10	0.03	27	0.06	0.01	18
Chlorine	wt.% d.b.	0.18	0.07	41	0.62	0.15	25
Ash composit	ion on ash 550	°C					
$Al_2O_3$	wt.% d.b.	1.6	0.4	24	0.48	0.38	78
CaO	wt.% d.b.	9.0	0.5	5.7	7.8	0.7	8.5
Fe <sub>2</sub> O <sub>3</sub>	wt.% d.b.	0.53	0.08	15	0.19	0.1	54
K <sub>2</sub> O	wt.% d.b.	23	3.0	13	21	1.6	7.5
MgO	wt.% d.b.	3.0	0.6	20	2.9	0.1	4.0
Na <sub>2</sub> O	wt.% d.b.	0.47	0.11	22	2.2	0.6	29
$P_2O_5$	wt.% d.b.	5.4	0.8	14	1.3	0.1	5.4
SO <sub>3</sub>	wt.% d.b.	5.8	1.0	14	2.7	0.2	8.9
$SiO_2$	wt.% d.b.	38	5.9	16	43	2.0	4.6
TiO <sub>2</sub>	wt.% d.b.	0.057	0.045	83	0.023	0.02	84
ZnO	wt.% d.b.	0.084	0.037	43	0.010	0.003	27
Sum		86.3			86.5		
Fusibility ten	nperatures						
Deformation	°C	900	73	8.2	810	22	2.8
Hemisphere	°C	1090	33	3	1030	18	1.7
Fluid	°C	1170	57	4.9	1090	16	1.5

Table 4: Characterisation of the barley and tall wheatgrass baled bioUnitTall wheatgrass (n=6)Barley straw (n=7)

d.b.: dry basis; GCV: gross calorific value at constant volume; NCV: net calorific value at constant pressure; ND: not determined

Loaching	troatmont	Tall wh	eatgrass	Barley straw					
Leaching	lieatment	DT	FT	DT	FT				
	Initial	900	1170	810	1090				
	1 hour	950	1280	940	1260				
	1 hour+ stirring	910	1140	880	1240				
inmersion	5 hours	1070	1300	1120	1280				
	1 day	1160	1360	1220	1400				
	2 days	1320	> 1450	1280	> 1450				
Irrigation	900 L	830	1150	870	1150				
ingation	1800 L	ND	ND	1160	1250				
DT. dafa		LT. fl.		ND. and Jat	a mana i mana al				

**Table 5:** Ash melting characteristic temperatures (according to ISO 21404:2020)

 depending on the leaching treatment.

DT: deformation temperature; FT: fluid temperature; ND: not determined

<u>\*</u>



ImmersionIrrigationFigure 1: Scheme of the two different bale leaching treatments



Figure 2: Schematics of the corrosion test using a tubular furnace.



Figure 3: S, Cl, and K concentrations on the baled tall wheatgrass biomass depending on the leaching treatment.



Figure 4: S, Cl, and K concentrations on the baled barley straw biomass depending on the leaching treatment.



Figure 5: Ash inorganic compounds of tall wheatgrass bales depending on the leaching treatment (XRD analysis)



Figure 6: Ash inorganic compounds of barley straw bales depending on the leaching treatment (XRD analysis)



Figure 7: Earth-alkaline to alkaline oxides ratio (empirical index I) depending on the leaching treatment.



Figure 8: Metal corrosion (SEM analysis) depending on the leaching treatment.



Figure 9: Ash/steel interfaces before and after 1-day bale immersions (SEM images).



Figure 10: High-density bales natural drying after being subjected to different leaching treatments.